



# Role of $\text{Ni}^{2+}$ ions in $\text{TiO}_2$ and $\text{Pt}/\text{TiO}_2$ photocatalysis for phenol degradation in aqueous suspensions

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## ABSTRACT

Anatase  $\text{TiO}_2$  is the most studied environmental photocatalyst, but its efficiency is still not high enough to enable application. Herein we report a positive effect of  $\text{Ni}^{2+}$  ions on the photocatalytic reaction in neutral aqueous solution. On addition of 1 mM  $\text{Ni}(\text{ClO}_4)_2$ , the rates of phenol oxidation on  $\text{TiO}_2$  and 0.5 wt.%  $\text{Pt}/\text{TiO}_2$  increased by factors of 2.1 and 8.0, respectively. Meanwhile, the formation of organic intermediate was reduced, and the removal of total organic content was enhanced. Interestingly,  $\text{Ni}^{2+}$  concentration did not change with irradiation time. Moreover, a maximum rate of phenol degradation was observed with  $\text{Ni}^{2+}$  ions, occupying half the surface of  $\text{TiO}_2$  or  $\text{Pt}/\text{TiO}_2$ , where  $\text{Ni}(\text{OH})^+$  was more active than  $\text{Ni}(\text{OH})_2$ . X-ray photoelectron spectroscopy identified a  $\text{Ni}^{3+}$  species produced in absence of phenol. A (photo)electrochemical study revealed that  $\text{Ni}^{2+}$  oxidation was favored over  $\text{H}_2\text{O}$  oxidation, whereas  $\text{Ni}^{2+}$  and Pt inhibited and facilitated  $\text{O}_2$  reduction, respectively. It is proposed that the adsorbed  $\text{Ni}(\text{OH})^+$  on  $\text{TiO}_2$  is oxidized to  $\text{Ni}^{3+}$  by the photogenerated holes of  $\text{TiO}_2$ , followed by regeneration through phenol oxidation. The  $\text{Ni}^{2+}$ -catalyzed organic oxidation then promotes the Pt-catalyzed  $\text{O}_2$  reduction, and vice versa, resulting into the greatly improved efficiency of charge separation. This work highlights the possibility of  $\text{Ni}^{2+}$  ions as the hole mediator of organic oxidation over a semiconductor.

## 1. Introduction

Semiconductor photocatalysis for energy and environmental use has been extensively studied over the past forth decades. The most applied reactions include water splitting to  $\text{H}_2$  and  $\text{O}_2$  [1],  $\text{CO}_2$  reduction to organics with  $\text{H}_2\text{O}$  as reductant [2], and organic degradation to  $\text{CO}_2$  with  $\text{O}_2$  as oxidant [3]. These reactions have a common feature. Semiconductor first absorbs photons, and generates pairs of electrons and holes in the conduction and valence bands, respectively. Then these charge carriers migrate onto the surface, reacting with sorbates. Due to the fast recombination of charge carriers, however, the quantum yield of a target reaction is usually low. Although new materials need to be developed [4], the central issue of photocatalysis is how to accelerate the interfacial charge transfer through a simple and economic way.

Among various photocatalysts,  $\text{TiO}_2$  is the most studied for environmental use, mainly due to its superior band structures [5,6]. In neutral aqueous solution, the conduction band edge potential of anatase  $\text{TiO}_2$  are  $-0.53$  V versus normal hydrogen electrode (NHE) [7], which is more negative than the one-electron oxidation potential of  $\text{O}_2$  ( $-0.33$  V vs. NHE) [8,9]. The counterpart valence band edge potential is  $2.67$  V vs. NHE, which is more positive than the one-electron reduction potentials of many organic substrates [10]. To improve the

quantum efficiency of organic degradation, the surface of  $\text{TiO}_2$  is usually modified with a cocatalyst [11]. For example, noble metals such as Pt can catalyze the electron reduction of  $\text{O}_2$ , and hence promote the hole oxidation of organics [12,13]. In acidic aqueous solution, both  $\text{Fe}^{3+}$  and  $\text{Cu}^{2+}$  ions can capture the electrons of  $\text{TiO}_2$ , improving the efficiency of charge separation for organic oxidation [14,15]. But the reoxidation of  $\text{Fe}^{2+}$  to  $\text{Fe}^{3+}$  by  $\text{O}_2$  is slow, whereas the reoxidation of  $\text{Cu}^+$  to  $\text{Cu}^{2+}$  by  $\text{O}_2$  is efficient only at low concentration. In contrary,  $\text{Co}^{2+}$  ions can capture the holes of  $\text{TiO}_2$  to form a highly reactive  $\text{Co}^{\text{III/IV}}$  for organic oxidation. Without precipitating reagent such as phosphate, however, the  $\text{Co}^{\text{III/IV}}$  species deactivate to  $\text{Co}^{2+}$ , making  $\text{Co}^{2+}$  in null cycle, and consequently slowing down the  $\text{TiO}_2$ -photocatalyzed reaction [16].

In this work, the effect of  $\text{Ni}^{2+}$  ions on the  $\text{TiO}_2$ -photocatalyzed degradation of organics is examined. The present work focus on the surface adsorbed  $\text{Ni}^{2+}$  on  $\text{TiO}_2$ , which is different from the bulk doped  $\text{Ni}^{2+}$  in  $\text{TiO}_2$  [17]. There is also a debate about  $\text{Ni}^{2+}$  reduction or oxidation occurring or not on the irradiated  $\text{TiO}_2$  [18]. Phenol represents a class of aromatic pollutants, and it is often used as a model substrate to assess the catalyst photoactivity, because it is colorless, and its degradation under light at wavelengths longer than  $320$  nm result only from the  $\text{TiO}_2$ -photocatalyzed reaction [19]. To accelerate the

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electron transfer,  $\text{TiO}_2$  was also deposited with 0.5% of Pt nanoparticles. On addition of 1 mM  $\text{Ni}(\text{ClO}_4)_2$ , interestingly, the rates of phenol degradation on  $\text{TiO}_2$  and Pt/ $\text{TiO}_2$  increased by factors of 2.1 and 8.0, respectively. But  $\text{Ni}^{2+}$  concentration in aqueous phase remained unchanged with time, either in the absence or presence of phenol. To understand the role of  $\text{Ni}^{2+}$ , several experiments were carried out, including the change of Ni valence state on solid, the electrochemical reduction of  $\text{O}_2$ , and the photoelectrochemical oxidation of phenol and water on a film electrode.

## 2. Experimental

### 2.1. Materials

Nickel perchlorate and anatase  $\text{TiO}_2$  were purchased from Sigma-Aldrich. Dimethylglyoxime and ammonium peroxydisulfate were purchased from Aladdin, and others from Shanghai Chemicals. Pt/ $\text{TiO}_2$  was made by a photochemical method. A suspension (50 mL) containing 18 g/L  $\text{TiO}_2$ , 2 M  $\text{CH}_3\text{OH}$ , and a certain amount of  $\text{H}_2\text{PtCl}_6$  was irradiated with a 300 W mercury lamp (300 W) for 3 h. Then the solid was collected, and washed several times with water, and dried at 80 °C. The amount of Pt in Pt/ $\text{TiO}_2$  was calculated to be 0.5% weight percent.

### 2.2. Characterization

X-ray diffraction (XRD) patterns were recorded on a D/max-2550/PC diffractometer (Rigaku). Transmission electron microscopy (TEM) was recorded on a JEM-2100 F field-emission instrument. Diffuse reflectance spectra (DRS) were recorded on a Shimadzu UV-2600 with  $\text{BaSO}_4$  as reference. Reflectance ( $R$ ) was transferred into Kubelka–Munk absorbance ( $F_R$ ), equal to  $(1-R)^2/(2R)$ . Adsorption isotherms of  $\text{N}_2$  on solid were measured at 77 K on a Micromeritics ASAP2020 apparatus. X-ray photoelectron spectroscopy (XPS) spectra were recorded on a Kratos AXIS Ultra DLD spectrometer, followed by calibration with C 1s at 284.6 eV.

### 2.3. Photocatalysis and analysis

Experiments were carried out on a photochemical apparatus (Nanjing XuJiang Power Plant, China). Unless stated otherwise, reactions were performed at 25 °C under fixed initial conditions (0.43 mM phenol, 1.0 g/L  $\text{TiO}_2$ , 1.0 mM  $\text{Ni}^{2+}$ , and pH 7.0). The solution pH was adjusted with a dilute solution of  $\text{HClO}_4$  or  $\text{NaOH}$ . A suspension containing necessary components was stirred in the dark for 0.5 h, and then irradiated with a high pressure mercury lamp. The light intensity reaching the external surface of reactor was 4.50 mW  $\text{cm}^{-2}$ , measured with a UVA irradiance meter. At given intervals, 2.0 mL of the suspension was withdrawn, and filtered through a 0.22  $\mu\text{m}$  membrane. The filtrate was immediately analyzed by HPLC (high performance liquid chromatography) on a Dionex UltiMate 3000 (50%  $\text{CH}_3\text{OH}$  as eluent). TOC (total organic content) was analyzed on a TOC-V CPH TOC analyzer (Shimadzu).

### 2.4. Adsorption isotherm of $\text{Ni}^{2+}$

Measurement was carried out at 25 °C and in the dark. A suspension containing 2.0 g/L catalyst and a certain concentration of  $\text{Ni}^{2+}$  was stirred for 12 h, and then filtered through a 0.22  $\mu\text{m}$  membrane. Through a complex with dimethylglyoxime [20],  $\text{Ni}^{2+}$  ions were analyzed at 550 nm on an Agilent 8453 UV–vis spectrophotometer.

### 2.5. Recycle test of $\text{Ni}^{2+}$

Reaction was performed with Pt/ $\text{TiO}_2$  five times, in absence and presence of  $\text{Ni}^{2+}$ , respectively. For each run under similar conditions, two suspensions were prepared, which were A (100 mL, 2.0 g/L Pt/

$\text{TiO}_2$ , and pH 7.0), and B (100 mL, 2.0 g/L Pt/ $\text{TiO}_2$ , 1.0 mM  $\text{Ni}^{2+}$ , and pH 7.0). Each suspension was divided into two parts. One was continuously stirred in the dark. Another was mixed with 1.0 mL of phenol stock solution (11 mM), followed by stirring for 30 min, and irradiation with UV light. For A and B, total irradiation time was 70 and 15 min, respectively. During this period of time, five samples were taken out for analysis, and each of them was 2 mL. After that, the irradiated suspension was supplied with 10 mL of the un-irradiated suspension and 1 mL of phenol stock solution, followed by pH adjustment to initial pH 7.0. After stirred in the dark for 30 min, new suspension was irradiated again, and analyzed as described above.

### 2.6. (Photo)electrochemical measurement

A film working electrode was prepared by doctor blade method. A gel containing PVA (polyvinylalcohol) and  $\text{TiO}_2$  or Pt/ $\text{TiO}_2$  coated an indium-doped tin oxide (ITO) glass (Nippon Sheet Glass). To remove PVA, the film glass was annealed in air at 500 °C for 1 h. Then the glass was cut into several pieces. Each piece had an exposed area of 1.0  $\text{cm}^2$ , and other part was sealed by epoxy resin. Measurements were carried out on a CHI660E Electrochemical Station (Chenhua, Shanghai), using a platinum gauze as counter electrode, and an Ag/AgCl as reference electrode. The working electrode was illuminated with a 500 W Xenon lamp, from the electrode/electrolyte side. The electrolyte was 0.5 M  $\text{NaClO}_4$  (pH 7.0), and the scanning rate was 10 mV/s.

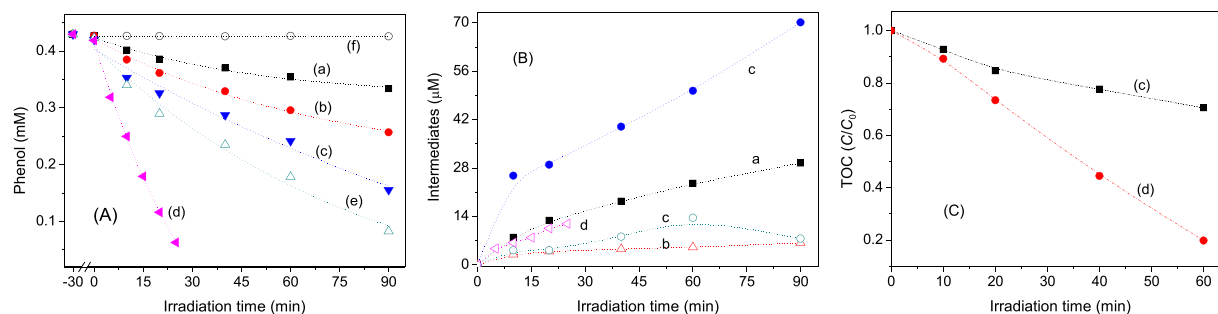
## 3. Results and discussion

### 3.1. Characterization

Solid was characterized with XRD, TEM, DRS,  $\text{N}_2$  adsorption, and XPS, and the results are shown in Figure S1 of the Supporting Materials. In brief,  $\text{TiO}_2$  and Pt/ $\text{TiO}_2$  had XRD patterns, well matching that of anatase  $\text{TiO}_2$  (PDF #65-5714), respectively. According to the strongest diffraction, two samples were similar in peak intensity, integrated area, and peak width at half height (Table S1). It means that  $\text{TiO}_2$  phase remains unchanged after Pt deposition. In the TEM images, the  $\text{TiO}_2$  particles were spherical-like, with a diameter of about 30 nm, either in absence or presence of Pt. In the adsorption–desorption isotherm of  $\text{N}_2$ , there was a hypothesis loop, indicative of mesopores present in samples. But the surface area, total pore volume, and average pore size of Pt/ $\text{TiO}_2$  were slightly smaller than those of  $\text{TiO}_2$ , respectively. It means that the solid pores are partially occupied or blocked by Pt fine particles. In the DRS spectrum, there was an intensive absorption at wavelengths shorter than 400 nm region, assigned to the band-to-band transition of  $\text{TiO}_2$ . Through a Tauc plot, the band gap energy ( $E_g$ ) for indirect transition was estimated to be 3.2 eV for both  $\text{TiO}_2$  and Pt/ $\text{TiO}_2$ . In the XPS spectrum of Pt 4f, there were two peaks at 69.73 and 73.79 eV, assigned to Pt 4f<sub>7/2</sub> and Pt 4f<sub>5/2</sub> of Pt<sup>0</sup>, respectively [21–24]. But the binding energy of Pt/ $\text{TiO}_2$  was 0.77 eV lower than that of Pt/C (71.5 eV), due to the shift of electron density from  $\text{TiO}_2$  to Pt [21,22]. These observations indicate that metallic Pt is deposited onto  $\text{TiO}_2$ , without obvious change in the crystal structure, crystallite size, and optical property of  $\text{TiO}_2$  phase.

### 3.2. Photocatalytic degradation of phenol

In this study,  $\text{Ni}(\text{ClO}_4)_2$  was used as  $\text{Ni}^{2+}$  source, because specific anions such as  $\text{F}^-$  and  $\text{SO}_4^{2-}$  have influence on  $\text{TiO}_2$  photocatalysis [19,25]. Fig. 1A shows the results of phenol degradation on  $\text{TiO}_2$  and Pt/ $\text{TiO}_2$ , measured in the absence and presence of 1 mM  $\text{Ni}^{2+}$ . Among the samples, the rates of phenol degradation were Pt/ $\text{TiO}_2$  +  $\text{Ni}^{2+}$  > P25 > Pt/ $\text{TiO}_2$  >  $\text{TiO}_2$  +  $\text{Ni}^{2+}$  >  $\text{TiO}_2$ , where P25 is a mixture of anatase and rutile, and it is often used as a reference photocatalyst. To confirm the effect of  $\text{Ni}^{2+}$ , several experiments under UV light were performed. First, in a homogeneous solution of  $\text{Ni}(\text{ClO}_4)_2$  or  $\text{NaClO}_4$ ,



**Fig. 1.** (A) Phenol degradation over (a) TiO<sub>2</sub>, (b) TiO<sub>2</sub> + Ni<sup>2+</sup>, (c) Pt/TiO<sub>2</sub>, (d) Pt/TiO<sub>2</sub> + Ni<sup>2+</sup>, (e) P25 TiO<sub>2</sub>, and (f) Ni<sup>2+</sup>. (B) The corresponding formation of hydroquinone (solid symbols) and benzoquinone (open symbols), and (C) change of total organic content (TOC).

there was negligible decrease in phenol concentration. That is, phenol was stable against UV light, even in the presence of Ni<sup>2+</sup>, Na<sup>+</sup>, and ClO<sub>4</sub><sup>−</sup>, respectively. Second, with TiO<sub>2</sub> or Pt/TiO<sub>2</sub> as photocatalyst, the addition of NaClO<sub>4</sub> did not change the rate of phenol degradation (Fig. S2A). Then the increased rate of phenol degradation on addition of Ni (ClO<sub>4</sub>)<sub>2</sub> is surely due to the positive of Ni<sup>2+</sup> cations, not ClO<sub>4</sub><sup>−</sup> anions. Third, on addition of 1 mM Ni(ClO<sub>4</sub>)<sub>2</sub>, Ni(NO<sub>3</sub>)<sub>2</sub>, and NiCl<sub>2</sub>, the rates of phenol degradation on TiO<sub>2</sub> were all increased, but Ni(ClO<sub>4</sub>)<sub>2</sub> was more active than other nickel salts (Fig. S2B). These observations indicate that Ni<sup>2+</sup> ions are beneficial to the photocatalytic degradation of phenol either on TiO<sub>2</sub> or Pt/TiO<sub>2</sub>.

All of the time profiles for phenol degradation well fit the pseudo-first-order rate equation, and the resulting apparent constants ( $k_{obs}$ ) are shown in Table 1. Such kinetics is often observed in TiO<sub>2</sub> photocatalysis, and attributed to the constant formation of reactive species under a fixed condition [26,27]. In the present work, phenol photolysis in aqueous solution and its adsorption on TiO<sub>2</sub> were both negligible. Then a large value of  $k_{obs}$  corresponds to a high concentration of reactive species produced. According to  $k_{obs}$ , one can say that the photocatalytic activity of Pt/TiO<sub>2</sub> is approximately 4 times that of TiO<sub>2</sub>. This is ascribed to the formation of a Schottky barrier allowing electron transfer from TiO<sub>2</sub> to Pt [12,28], and/or due to the thermal catalysis of Pt for O<sub>2</sub> reduction [21]. Both of them would inhibit electron recombination with the holes, and hence promote the hole oxidation of phenol. On the addition of Ni<sup>2+</sup>, strikingly, the photocatalytic activities of TiO<sub>2</sub> and Pt/TiO<sub>2</sub> increased by factors of 2.1 and 8.0, respectively. Moreover, the  $k_{obs}$  value of Pt/TiO<sub>2</sub> plus Ni<sup>2+</sup> was 4.7 times the sum of individual  $k_{obs}$  (TiO<sub>2</sub> plus Ni<sup>2+</sup>, and Pt/TiO<sub>2</sub>). These observations indicate that for phenol degradation, Ni<sup>2+</sup> ions are not only beneficial, but also have a cooperative effect with Pt, further improving the photocatalytic activity of TiO<sub>2</sub>.

During phenol degradation, benzoquinone (BQ) and hydroquinone (HQ) were identified as main intermediates, and the results are shown in Fig. 1B. In the absence or presence of Ni<sup>2+</sup>, total intermediates formed from Pt/TiO<sub>2</sub> were more than those produced from TiO<sub>2</sub> (Tables S2 and S3). This trend in intermediate production between two catalysts matches that in phenol degradation. However, with each catalyst, HQ was much more and less than BQ in the absence and presence of Ni<sup>2+</sup>, respectively, whereas the total intermediates measured with Ni<sup>2+</sup> were less than those obtained without Ni<sup>2+</sup>. These observations imply

that Ni<sup>2+</sup> not only helps HQ transformation to BQ, but also helps BQ degradation into some intermediates. The total organic content (TOC) was measured, and the result obtained with Pt/TiO<sub>2</sub> is shown in Fig. 1C. As irradiation time increased, TOC decreased gradually. But the decay of TOC with time in presence of Ni<sup>2+</sup> was faster than that in absence of Ni<sup>2+</sup>. For example, on addition of Ni<sup>2+</sup>, TOC removal yield at 60 min was greatly increased, from 30% to 80%. Moreover, on addition of Ni<sup>2+</sup>, the mole ratio of total intermediates produced to phenol consumed was also notably decreased, from 29–35% to 3.3–7.2% (Tables S2 and S3). These observations indicate that Ni<sup>2+</sup> ions are beneficial to both the photocatalytic oxidation and mineralization of phenol either on TiO<sub>2</sub> or Pt/TiO<sub>2</sub>.

### 3.3. Effect of variables

Fig. 2A shows the pH effect on the rate constants of phenol degradation over TiO<sub>2</sub> and Pt/TiO<sub>2</sub>. In absence of Ni<sup>2+</sup>, the rate of phenol degradation slightly increased with the suspension pH. In presence of 1 mM Ni<sup>2+</sup>, however, the reaction rates greatly changed with the initial pH. A maximum reaction rate was observed at approximately pH 7.0, either from TiO<sub>2</sub> or Pt/TiO<sub>2</sub>. Interestingly, the effect of Ni<sup>2+</sup> was great at initial pH 7–10, but it was negligible at initial pH 3–6. This is presumably due to the pH-dependent adsorption of Ni<sup>2+</sup> on solid. It is known that TiO<sub>2</sub> in aqueous solution is amphoteric [29]. Then the zeta potentials of solid in aqueous suspension at different pHs were measured, and the results are shown in Fig. S3A. At zero point of charge, the pH value (pH<sub>zpc</sub>) was determined to be 5.56 for TiO<sub>2</sub>, and 5.74 for Pt/TiO<sub>2</sub>, respectively. That is, at a pH lower than pH<sub>zpc</sub>, the surfaces of photocatalysts are positively charged, unfavorable to Ni<sup>2+</sup> adsorption. In fact, as the suspension pH increased from 3 to 7, the amount of Ni<sup>2+</sup> adsorbed on TiO<sub>2</sub> or Pt/TiO<sub>2</sub> in aqueous suspensions at pHs 3–7 greatly increased with pH (Fig. S3B). This is accordant with the reduced density of positive charges on solid. Since the pH-dependent change in the rate increase of phenol degradation matches that in the amount of Ni<sup>2+</sup> adsorption, it follows that the observed positive effect of Ni (ClO<sub>4</sub>)<sub>2</sub> is due to the Ni(II) species adsorbed on solid, rather than in solution. Moreover, there is also a distribution of Ni<sup>2+</sup> and NiOH<sup>+</sup> in aqueous solution (Ni<sup>2+</sup> + OH<sup>−</sup> = NiOH<sup>+</sup>,  $K_f = 4.47 \times 10^5$ ) [30]. A theoretical calculation with 1 mM Ni<sup>2+</sup> solution showed that the mole percent of Ni<sup>2+</sup> and NiOH<sup>+</sup> decreased and increased, respectively, as the solution pH was increased (Fig. S4A). For examples, in aqueous solution at pH 7, there are 96% Ni<sup>2+</sup> and 4% NiOH<sup>+</sup>. In other words, the adsorbed Ni(II) species on solid in a weakly aqueous solution would be NiOH<sup>+</sup>, which then takes part into the photocatalytic degradation of phenol.

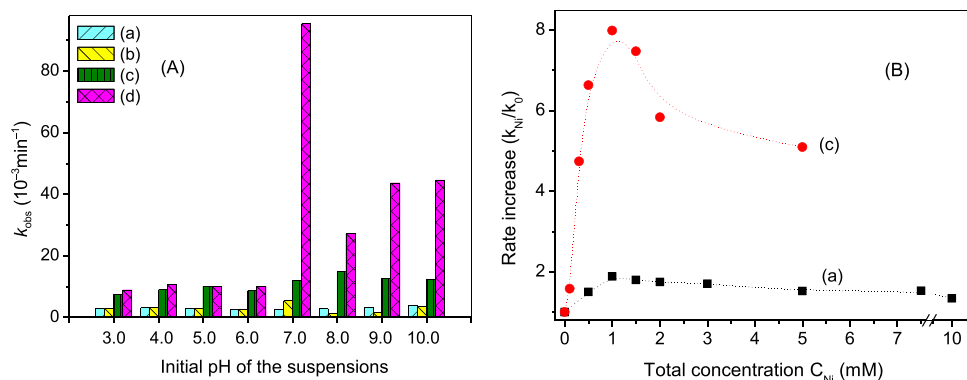
At a suspension pH higher than 7, however, a green precipitate was formed, due to the hydrolysis of Ni<sup>2+</sup> ions to Ni(OH)<sub>2</sub>. According to the solubility product of Ni(OH)<sub>2</sub> in water ( $K_{sp} = 7.94 \times 10^{-17}$ ) [30], a solution of 1 mM Ni<sup>2+</sup> would give Ni(OH)<sub>2</sub> at a pH equal to and higher than 7.5. Since phenol degradation at pH 7 was faster than those at pH 8–10, it follows that on the oxide surface, NiOH<sup>+</sup> ions are more active

**Table 1**  
Data of phenol degradation and Ni<sup>2+</sup> adsorption.

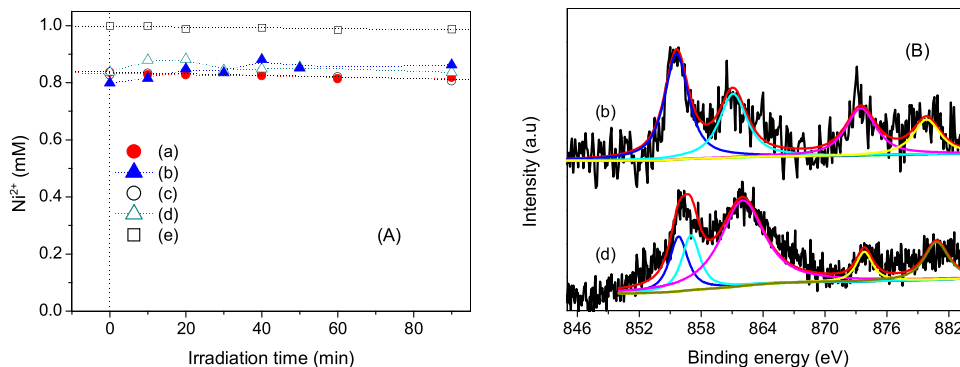
Sample	Phenol $k_{obs}$ (10 <sup>−3</sup> min <sup>−1</sup> ) <sup>a</sup>		Ni <sup>2+</sup> adsorption <sup>b</sup>	
	Without Ni <sup>2+</sup>	With 1 mM Ni <sup>2+</sup>	$q_m$ (mmol/g)	$K$ (mM <sup>−1</sup> )
TiO <sub>2</sub>	2.55	5.41	0.204	24.1
Pt/TiO <sub>2</sub>	10.2	73.3	0.197	25.8

<sup>a</sup>  $k_{obs}$ , apparent rate constant.

<sup>b</sup>  $q_m$ , the maximum amount of adsorption;  $K$ , adsorption constant.



**Fig. 2.** (A) Effect of initial pH on the rate constant of phenol degradation ( $k_{\text{obs}}$ ), measured in the presence of (a)  $\text{TiO}_2$ , (b)  $\text{TiO}_2 + 1.0 \text{ mM Ni}^{2+}$ , (c)  $\text{Pt/TiO}_2$ , and (d)  $\text{Pt/TiO}_2 + 1.0 \text{ mM Ni}^{2+}$ . (B) Effect of  $\text{Ni}^{2+}$  concentration on the rate increase ( $k_{\text{Ni}}/k_0$ ), where  $k_0$  and  $k_{\text{Ni}}$  represent the rate constants of phenol degradation at initial pH 7 in the absence and presence of  $\text{Ni}^{2+}$ , respectively.



**Fig. 3.** (A) Time profiles of  $\text{Ni}^{2+}$  concentration in aqueous phase, under different conditions of (a)  $\text{TiO}_2 + \text{Ni}^{2+} + \text{phenol}$ , (b)  $\text{Pt/TiO}_2 + \text{Ni}^{2+} + \text{phenol}$ , (c)  $\text{TiO}_2 + \text{Ni}^{2+}$ , (d)  $\text{Pt/TiO}_2 + \text{Ni}^{2+}$ , and (e)  $\text{Ni}^{2+}$ . (B) XPS spectrum of Ni 2p, recorded from samples (b) and (d), after irradiation for 90 min.

than  $\text{Ni(OH)}_2$ , in participation of the photocatalytic reaction. In a comparison with the loosely deposited  $\text{Ni(OH)}_2$  on  $\text{TiO}_2$ , the electrostatically adsorbed  $\text{NiOH}^+$  on  $\text{TiO}_2$  would have a stronger interaction, and hence a faster reaction with the irradiated  $\text{TiO}_2$ . The rate of phenol degradation at pH 8, lower than those at pHs 9 and 10, is probably due to incomplete deposition of a colloidal  $\text{Ni(OH)}_2$  onto solid, causing light scattering, reducing the number of photons reaching  $\text{TiO}_2$ , and consequently decreasing the rate of photoreaction.

Fig. 2B shows the effect of  $\text{Ni}^{2+}$  concentration ( $C_{\text{Ni}}$ ) at initial pH 7.0, where  $k_0$  and  $k_{\text{Ni}}$  represent the apparent rate constants of phenol degradation measured in absence and presence of  $\text{Ni}^{2+}$ , respectively. As  $C_{\text{Ni}}$  increased, the ratio of  $k_{\text{Ni}}$  to  $k_0$  increased initially, and then decreased. A maximum  $k_{\text{Ni}}/k_0$  was observed around 1.0 mM  $\text{Ni}^{2+}$ , either from  $\text{TiO}_2$  or  $\text{Pt/TiO}_2$ . However, the raise and decay of  $k_{\text{Ni}}/k_0$  with  $C_{\text{Ni}}$ , observed from  $\text{Pt/TiO}_2$ , were faster than those, obtained from  $\text{TiO}_2$ , respectively. That is, the effect of  $\text{Ni}^{2+}$  to  $\text{Pt/TiO}_2$  is larger than that to  $\text{TiO}_2$ . Note that in all of  $\text{Ni}^{2+}$  solutions at pH 7, there are always 96%  $\text{Ni}^{2+}$  and 4%  $\text{NiOH}^+$  (Figure S4B). That is, the observed effect of  $C_{\text{Ni}}$  is not due to the change in the distribution of Ni species.

To understand the different effect of  $C_{\text{Ni}}$ , the adsorption isotherms of  $\text{Ni}^{2+}$  on solid in aqueous suspension at pH 7.0 were measured, and the results are shown in Fig. S3C. As the equilibrium concentration of  $\text{Ni}^{2+}$  in aqueous phase ( $C_{\text{eq}}$ ) increased, the amount of  $\text{Ni}^{2+}$  adsorbed on solid ( $q$ ) increased toward a limit. The isotherm well fit the Langmuir adsorption equation,  $q/q_m = KC_{\text{eq}}/(1 + KC_{\text{eq}})$ , where  $q_m$  is the maximum amount of adsorption, and  $K$  is adsorption constant. The  $q_m$  and  $K$  values of  $\text{Pt/TiO}_2$  are slightly smaller and larger than those of  $\text{TiO}_2$ , respectively (Table 1). The former matches the change in solid surface area and pore volume (Table S1). The latter is probably indicative of a special interaction between  $\text{Ni}^{2+}$  and Pt. In the region of low  $C_{\text{Ni}}$ , there was a linear correlation between  $k_{\text{Ni}}/k_0$  and  $q$  (Fig. 2B and S3B). This is in agreement with the above proposal that the positive effect of  $\text{Ni}^{2+}$  is due to the adsorbed  $\text{NiOH}^+$  on solid. Interestingly, a maximum  $k_{\text{Ni}}/k_0$  was observed at 1 mM  $\text{Ni}^{2+}$ , which occupied 50% of  $\text{TiO}_2$  surface, and

58% of  $\text{Pt/TiO}_2$ , respectively. Beyond the optimum surface coverage,  $k_{\text{Ni}}/k_0$  became to decrease with  $C_{\text{Ni}}$  for  $\text{TiO}_2$ . According to  $K_{\text{sp}}$ , in a solution at pH 7,  $\text{Ni(OH)}_2$  is formed only when  $C_{\text{Ni}}$  is higher than 7.94 mM. Then the decreased rate of phenol degradation at 1–5 mM  $\text{Ni}^{2+}$  is due either to the green solution of  $\text{Ni}^{2+}$ , reducing the number of photons reaching photocatalyst, or to the adsorbed Ni species on solid, decreasing the amount of  $\text{O}_2$  adsorption on photocatalyst. Both of them would slow down the photocatalytic reaction for phenol degradation. However, in the range of  $C_{\text{Ni}}$  examined (0.1–5 mM), the rates of phenol degradation on  $\text{TiO}_2$  and  $\text{Pt/TiO}_2$  are all larger than those without  $\text{Ni}^{2+}$ , respectively. These observations indicate that the adsorbed  $\text{NiOH}^+$  species on  $\text{TiO}_2$  and  $\text{Pt/TiO}_2$  are indeed beneficial to the photocatalytic degradation of phenol.

### 3.4. Recycle behavior of $\text{Ni}^{2+}$ ions

During phenol degradation,  $\text{Ni}^{2+}$  concentration in aqueous phase was monitored, and the results are shown in Fig. 3A. In the dark, about 18% of  $\text{Ni}^{2+}$  ions were adsorbed on solid. Under UV light,  $\text{Ni}^{2+}$  concentration did not change with time, on the average. At the same time, however, the suspension changed its color from white to pale brown, as observed notably from  $\text{Pt/TiO}_2$  (Fig. S5A). Such color change is due to the formation of yellowish BQ from phenol degradation (Fig. 1B), and/or due to the change of solid chemical state. To clarify it,  $\text{Pt/TiO}_2$  was collected, dried, and analyzed by XPS. In the spectrum of Pt 4f, no difference was observed between the dark and irradiated samples (Fig. S1C and S1D). In the spectrum of Ni 2p, there were two peaks at 855.3 and 873.1 eV (Fig. 3B), assigned to  $\text{Ni } 2p_{3/2}$  and  $\text{Ni } 2p_{1/2}$ , respectively. But there was also no difference between the dark and irradiated samples (Fig. S6A and S6B). Moreover, both the peak position and separation (17.8 eV) of Ni 2p were similar to those for a  $\text{Ni(OH)}_2$  powder [31], and a  $\text{Ni(OH)}_2$  thick film on  $\text{TiO}_2$  and  $\text{Fe}_2\text{O}_3$  [32–35]. These observations indicate that during phenol degradation,  $\text{Ni}^{2+}$  ions remain unchanged, either in concentration or chemical state.



A question arose regarding how  $\text{Ni}^{2+}$  ions were beneficial to phenol degradation. To examine the fate of  $\text{Ni}^{2+}$ , experiment with Pt/ $\text{TiO}_2$  and 1 mM  $\text{Ni}^{2+}$  was carried out in the absence of phenol. Under UV light,  $\text{Ni}^{2+}$  concentration in aqueous phase remained unchanged with time (Fig. 3A, open symbols). At the same time, the suspension changed its color from white to brown. But such color change in absence of phenol (Fig. S5B) was faster than that in presence of phenol (Fig. S5A). Control experiments with Pt/ $\text{TiO}_2$  (Fig. S5C), and  $\text{TiO}_2$  plus  $\text{Ni}^{2+}$  (Fig. S5D) showed negligible and slow color changes, respectively. These observations imply that a new Ni species is formed on the irradiated catalyst, which is reactive to phenol, but not to water. To identify it, Pt/ $\text{TiO}_2$  was collected, dried, and analyzed by XPS. In the spectrum of Ni 2p, there was an obvious difference between the dark and irradiated samples (Fig. S6C and S6D). After light irradiation, the Ni 2p<sub>3/2</sub> peak was shifted by 0.8 eV, together with a 1.0 eV increase in the peak width at half height. Since the peak was asymmetric, it was deconvoluted into two peaks at 855.8 and 857.0 eV, respectively (Fig. 3B). The former matches that for Ni 2p<sub>3/2</sub> of  $\text{Ni}^{2+}$ . The latter is assigned to Ni 2p<sub>3/2</sub> of  $\text{Ni}^{3+}$ , as observed from a photoelectrochemically oxidized  $\text{Ni}(\text{OH})_2$  thick film on  $\text{Fe}_2\text{O}_3$  [34,35]. But the XPS peaks due to NiO (854.3 eV) and  $\text{Ni}^0$  (852.7 eV) were absent in the present sample [32,33,36,37]. These observations indicate that the adsorbed  $\text{NiOH}^+$  ions on solid under UV light are oxidized to a  $\text{Ni}^{3+}$  species, mostly in the form of  $\text{NiOOH}$ . Then  $\text{NiOOH}$  oxidizes phenol to regenerate  $\text{Ni}^{2+}$ , followed by the re-adsorption of  $\text{Ni}^{2+}$  onto solid. Since  $\text{NiOOH}$  is surface bound, and not reactive to  $\text{H}_2\text{O}$ ,  $\text{Ni}^{2+}$  concentration in aqueous phase remains unchanged with irradiation time, either in the absence or presence of phenol (Fig. 3A).

To examine the recycle behavior of  $\text{Ni}^{2+}$ , phenol degradation on Pt/ $\text{TiO}_2$  was repeated five times, and the results are shown in Fig. 4A. For the reactions in the absence and presence of 1 mM  $\text{Ni}^{2+}$ , total irradiation time was fixed to be 70 and 15 min, respectively. In each run, phenol degradation with  $\text{Ni}^{2+}$  was always faster than that without  $\text{Ni}^{2+}$ . But from one run to another, the reaction rate gradually decreased, either in the absence or presence of  $\text{Ni}^{2+}$  (Table S4). This is mainly due to organic intermediates formed and accumulated in the reactor, competing with phenol for reactive species. However, from the first run to last, the ratio of  $k_{\text{obs}}$  in presence of  $\text{Ni}^{2+}$  to that in absence of  $\text{Ni}^{2+}$  remained nearly unchanged. It implies that  $\text{Ni}^{2+}$  ions are stable during the five repeated test. Fig. 4B and C show the time profiles for HQ and BQ formation. From one run to another, the intermediates gradually accumulated in the reactor. After the last run, total intermediates, formed in absence and presence of  $\text{Ni}^{2+}$ , were 60 and 20  $\mu\text{M}$ , respectively. But they were all much less than total phenol disappeared (1.55 mM without  $\text{Ni}^{2+}$ , and 1.30 mM with  $\text{Ni}^{2+}$ ). These observations indicate that  $\text{Ni}^{2+}$  ions can be repeatedly used for phenol degradation.

### 3.5. (Photo)electrochemical measurement

To further understand the effect of  $\text{Ni}^{2+}$ , the electron reduction of  $\text{O}_2$  and the hole oxidation of phenol were examined independently.

Fig. 5 shows the electrochemical reduction of  $\text{O}_2$  in a 0.5 M  $\text{NaClO}_4$  at pH 7. As the applied potential swept negatively, the dark current increased. At given potential, the electrode current under air was much larger than that under  $\text{N}_2$ . This is indicative of  $\text{O}_2$  reduction being dominant process under air. By using a literature method [38], the onset potential of  $\text{O}_2$  reduction was estimated, which was  $-0.152$  and  $-0.142$  V vs.  $\text{AgCl}/\text{Ag}$  for  $\text{TiO}_2$  and Pt/ $\text{TiO}_2$ , respectively. On addition of 1 mM  $\text{Ni}^{2+}$ , the onset potentials of  $\text{TiO}_2$  and Pt/ $\text{TiO}_2$  negatively shifted to be  $-0.164$  and  $-0.156$  V vs.  $\text{AgCl}/\text{Ag}$ , respectively. Meanwhile, the electrode current was decreased, the degree of which was larger with  $\text{TiO}_2$  than that with Pt/ $\text{TiO}_2$ . Regardless of  $\text{Ni}^{2+}$  present or not, moreover, the current of  $\text{O}_2$  reduction on Pt/ $\text{TiO}_2$  was always much larger than that on  $\text{TiO}_2$ . A more positive onset potential, and a larger current correspond to an easier reduction of  $\text{O}_2$ . These observations indicate that Pt efficiently catalyzes  $\text{O}_2$  reduction [21], but  $\text{Ni}^{2+}$  ions are negative to  $\text{O}_2$  reduction, either on  $\text{TiO}_2$  or Pt/ $\text{TiO}_2$ . The inhibitory effect of  $\text{Ni}^{2+}$  on  $\text{O}_2$  reduction is ascribed to the reduced amount of  $\text{O}_2$  adsorption on the electrode, as proposed above for the observed decrease in the rate of phenol degradation with  $C_{\text{Ni}}$  at high  $C_{\text{Ni}}$  (Fig. 2B). Since Pt acts as electron traps and promotes  $\text{O}_2$  reduction, such detrimental effect of  $\text{Ni}^{2+}$  for  $\text{O}_2$  reduction on Pt/ $\text{TiO}_2$  electrode is less serious than that on  $\text{TiO}_2$  electrode.

Fig. 6 shows the photoelectrochemical oxidation of water and phenol on  $\text{TiO}_2$  and Pt/ $\text{TiO}_2$  film electrodes in a 0.5 M  $\text{NaClO}_4$  at pH 7, measured in absence and presence of 1 mM  $\text{Ni}^{2+}$ . Before experiment, the electrolyte was purged with  $\text{N}_2$ , and then the cell was closed. In all cases, the electrode photocurrent increased, as the applied potential swept positively. This is due to the improved efficiency of charge separation for the hole reaction on the working electrode. But the effects of  $\text{Ni}^{2+}$  on water and phenol oxidation are different and complicated. In absence of phenol, both  $\text{TiO}_2$  and Pt/ $\text{TiO}_2$  electrodes showed a negative shift in onset potential on addition of  $\text{Ni}^{2+}$ . This is due to  $\text{Ni}^{2+}$  oxidation easier than water oxidation, as observed from a  $\text{Ni}(\text{OH})_2$  film on  $\text{Fe}_2\text{O}_3$  photoanode [35]. To confirm it, the dark voltage-current curves for  $\text{TiO}_2$  and Pt/ $\text{TiO}_2$  electrodes were measured under  $\text{N}_2$  in a 0.5 M  $\text{NaClO}_4$  at pH 7, and the results are shown in Fig. S7A and 7B. With each electrode, the potential onset of water oxidation was 0.95 V vs.  $\text{AgCl}/\text{Ag}$ . In presence of 1 mM  $\text{Ni}^{2+}$ , the onset potential shifted to 0.80 V vs.  $\text{AgCl}/\text{Ag}$ . Such potential at pH 7 was reasonably more positive than a cyclic voltammetry potential of 0.70 V vs.  $\text{AgCl}/\text{Ag}$ , obtained at pH 10 from a  $\text{Ni}(\text{OH})_2$  deposited Pt/ $\text{TiO}_2$  electrode (Fig. S7C). These observations confirm that the  $\text{Ni}^{2+}$ -caused negative shift in onset potential is due to  $\text{Ni}^{2+}$  oxidation, rather than  $\text{H}_2\text{O}$  oxidation. On the other hand, the photocurrents of  $\text{TiO}_2$  and Pt/ $\text{TiO}_2$  electrodes on addition of  $\text{Ni}^{2+}$  were decreased and increased, respectively. The decreased photocurrent of  $\text{TiO}_2$  is due to brown  $\text{NiOOH}$  produced, which absorbs light, and reduce the number of photons reaching  $\text{TiO}_2$ . The increased photocurrent of Pt/ $\text{TiO}_2$  is due to the hole oxidation of  $\text{H}_2\text{O}_2$ , formed from the Pt catalyzed reduction of  $\text{O}_2$  (note that  $\text{O}_2$  results from water oxidation). In absence or presence of  $\text{Ni}^{2+}$ , moreover, the photocurrent of Pt/ $\text{TiO}_2$  was much smaller than that of  $\text{TiO}_2$ . This is

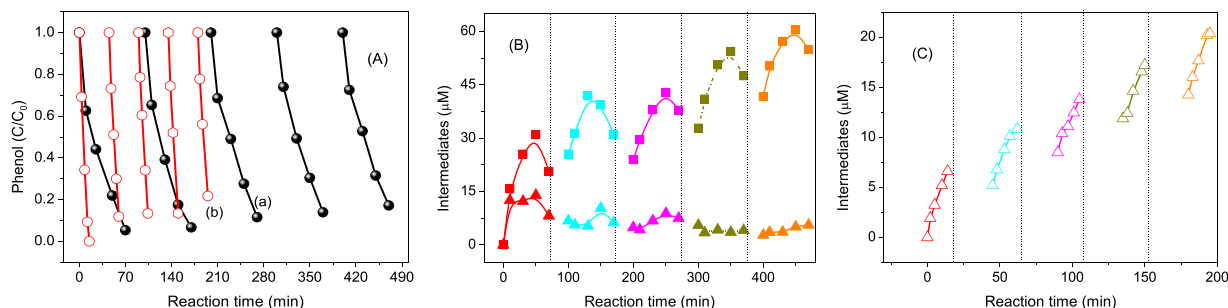
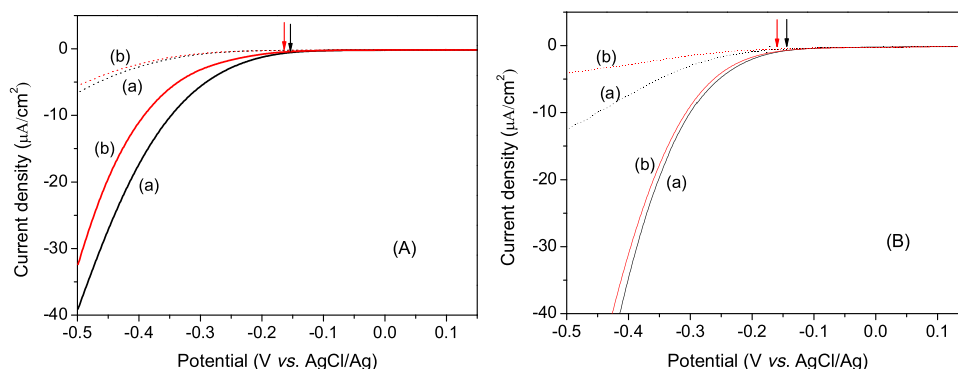


Fig. 4. (A) Phenol degradation on Pt/ $\text{TiO}_2$  (a) in absence, and (b) presence of  $\text{Ni}^{2+}$ . The formation of hydroquinone (square symbols) and benzoquinone (triangle symbols), (B) without, and (b) with  $\text{Ni}^{2+}$ . Initial conditions of each run: 2.0 g/L catalyst, 0.215 mM phenol, 1 mM  $\text{Ni}^{2+}$ , and pH 7.



**Fig. 5.** Dark current-voltage curves of (A)  $\text{TiO}_2$ , and (B)  $\text{Pt/TiO}_2$  film electrodes, measured under  $\text{N}_2$  (dotted lines) and air (solid lines), (a) in the absence, and (b) presence of  $1.0 \text{ mM Ni}^{2+}$ .

ascribed to the electrons trapping by Pt, slowing down the electron flow to generate photocurrent [13].

In the presence of phenol, the onset potentials of  $\text{TiO}_2$  and  $\text{Pt/TiO}_2$  were also negatively shifted, on the addition of  $\text{Ni}^{2+}$ . This is ascribed to the hole oxidation of  $\text{Ni}^{2+}$  easier than phenol oxidation. But at an external potential more positive than approximately  $0 \text{ V vs. AgCl/Ag}$ , both of the electrodes showed an increased photocurrent, different from those observed in the absence of phenol. Accordingly, the increased photocurrent of each electrode on addition of  $\text{Ni}^{2+}$  is due to phenol oxidation by  $\text{NiOOH}$ . At an applied potential more negative than approximately  $0 \text{ V vs. AgCl/Ag}$ , the electron reduction of  $\text{NiOOH}$  to  $\text{Ni}^{2+}$  might also occur, with a rate higher than that of phenol oxidation by  $\text{NiOOH}$ . That is, there is a competition between the formation and reduction of  $\text{NiOOH}$  on the electrode. As a result, the increased photocurrent due to phenol oxidation by  $\text{NiOOH}$  is observed only after the photoelectrons are efficiently removed. These observations indicate that  $\text{NiOOH}$  formed on irradiated  $\text{TiO}_2$  or  $\text{Pt/TiO}_2$  is reactive to phenol, but not to water.

### 3.6. Possible mechanism

To understand the effect of  $\text{Ni}^{2+}$ , one needs to know the band gap energy ( $E_g$ ), conduction band edge potential ( $E_{cb}$ ), and valence band edge potential ( $E_{vb}$ ) of semiconductor, where  $E_g = E_{vb} - E_{cb}$ . For n-type semiconductor such as  $\text{TiO}_2$ ,  $E_{cb}$  is close to  $E_{fb}$ , the flat band potential. In this study,  $E_{cb}$  was measured through a Mott–Schottky plot, and it was then used to calculate  $E_{cb}$  and  $E_{vb}$  (see details in Fig. S8). Experiment was performed in a  $0.5 \text{ M NaClO}_4$  at pH 7, and the results are shown in Table 2. In the absence of  $\text{Ni}^{2+}$ , both  $\text{TiO}_2$  and  $\text{Pt/TiO}_2$  have  $E_{cb}$  more negative than the potential of  $\text{O}_2/\text{O}_2^-$  couple ( $-0.33 \text{ V vs NHE}$ ) [8,9], whereas their  $E_{vb}$  are more positive than the potential of phenol<sup>+</sup>/phenol couple ( $1.03 \text{ V vs. NHE}$ ) [10]. Then in thermodynamics the electron reduction of  $\text{O}_2$  and the hole oxidation of phenol are all possible. On the addition of  $\text{Ni}^{2+}$ , both  $\text{TiO}_2$  and  $\text{Pt/TiO}_2$  showed

**Table 2**

Band edge potentials for semiconductors in a  $0.5 \text{ M NaClO}_4$  at pH 7<sup>a</sup>.

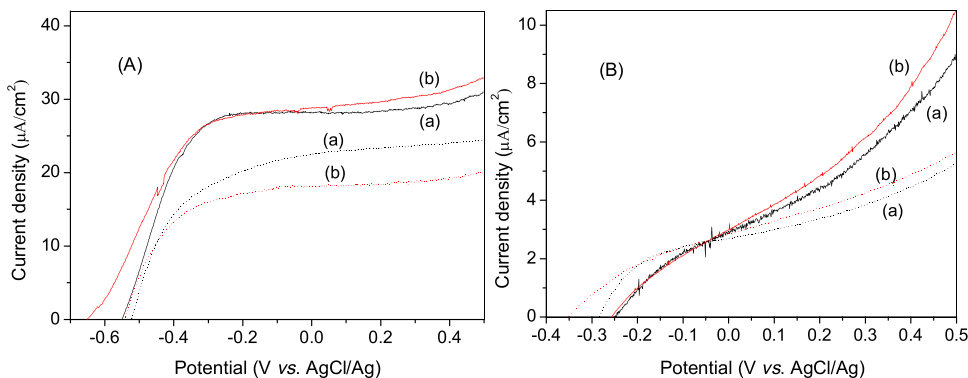
Sample	$E_{fb}$ (V, NHE)	$E_{cb}$ (V, NHE)	$E_{vb}$ (V, NHE)
$\text{TiO}_2$	$-0.34$ ( $-0.42$ )	$-0.47$ ( $-0.52$ )	$2.73$ ( $2.68$ )
$\text{Pt/TiO}_2$	$-0.42$ ( $-0.59$ )	$-0.55$ ( $-0.71$ )	$2.65$ ( $2.49$ )

<sup>a</sup> Bracket data represent ones measured in presence of  $1 \text{ mM Ni(ClO}_4)_2$ .  $E_{fb}$ , flat band potential;  $E_{cb}$ , conduction band edge potential;  $E_{vb}$ , valence band edge potential. Details in experiment and calculation method are shown in Fig. S8.

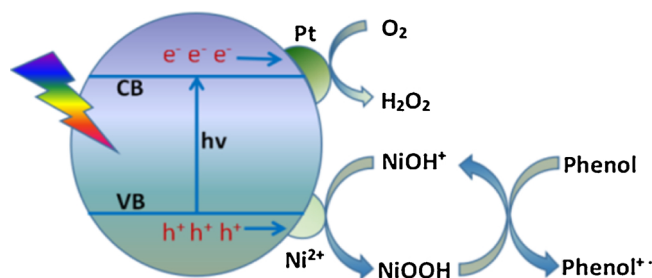
a negative shift in band edge potential. But their electrons to reduce  $\text{O}_2$  and their holes to oxidize phenol are still allowed in thermodynamics. As a result, phenol degradation was observed, either in the absence or presence of  $\text{Ni}^{2+}$  (Fig. 1A).

In aqueous solution at pH 7, the electrochemical oxidation of  $\text{Ni}^{2+}$  occurred around  $0.80 \text{ V vs. AgCl/Ag}$ , or  $1.0 \text{ V vs. NHE}$  (Fig. S7). This potential is less positive than  $E_{vb}$  of  $\text{TiO}_2$  and  $\text{Pt/TiO}_2$ , respectively. Then in thermodynamics, the hole oxidation of  $\text{Ni}^{2+}$  on two photocatalysts are both possible. In fact, the photocatalytic formation of  $\text{NiOOH}$  on  $\text{Pt/TiO}_2$  in absence of phenol was observed by XPS (Fig. 3B). On the other hand, the potential for water oxidation to  $\text{O}_2$  is  $0.82 \text{ V vs. NHE}$ . This potential is less positive than  $E_{vb}$  of two semiconductors ( $2.7 \text{ V vs. NHE}$ ), as well as the potential for  $\text{Ni}^{2+}$  oxidation ( $1.0 \text{ V vs. NHE}$ ). In practice, water oxidation by the photoholes of  $\text{TiO}_2$  and  $\text{Pt/TiO}_2$  were both observed (Fig. 6), but water oxidation by  $\text{NiOOH}$  was not observed (Fig. 3B). The non-reactivity of  $\text{NiOOH}$  toward water oxidation is also inferred from literature work. A photoelectrochemically oxidized  $\text{Ni(OH)}_2$  thick film on  $\text{TiO}_2$  can persist its brown color in air for several days [32,39]. Nevertheless, the larger photocurrent of  $\text{Ni(OH)}_2/\text{Fe}_2\text{O}_3$  than that of  $\text{Fe}_2\text{O}_3$  is ascribed to the oxidation of water by  $\text{Ni}^{IV}$  species, rather than by  $\text{Ni}^{III}$  [34,35].

The  $\text{NiOOH}$  produced from the hole oxidation of  $\text{NiOH}^+$  can oxidize phenol, HQ and BQ (Figs. 1 and 4). Then  $\text{Ni}^{2+}$  ions are regenerated,



**Fig. 6.** Photoelectrochemical oxidation of water (dotted lines) and  $0.43 \text{ mM}$  phenol (solid lines) on (A)  $\text{TiO}_2$ , and (B)  $\text{Pt/TiO}_2$  film electrodes under  $\text{N}_2$  in  $0.5 \text{ M NaClO}_4$ , measured (a) without  $\text{Ni}^{2+}$ , and (b) with  $1 \text{ mM Ni}^{2+}$ . Note that the same electrode was used for (a) and (b), but different one was used for phenol and water oxidation, respectively.



**Scheme 1.** A possible mechanism for the mutual effect of Pt and NiOH<sup>+</sup> on TiO<sub>2</sub> photocatalysis.

followed by subsequent hole oxidation to NiOOH. As a result, Ni<sup>2+</sup> concentration in solution remains unchanged during phenol degradation (Fig. 3A). In aqueous solution, Ni<sup>2+</sup> ions highly adsorb onto TiO<sub>2</sub> (0.2 mmol/g, Table 1), as compared with phenol (less than 0.01 mmol/g). Then, more photoholes of TiO<sub>2</sub> are consumed by Ni<sup>2+</sup> ions, and more photoelectrons of TiO<sub>2</sub> are available to O<sub>2</sub> reduction. As a result, the efficiency of charge separation is improved, and the rate of phenol oxidation is increased.

In TiO<sub>2</sub> photocatalysis, the electrons and holes are photogenerated and consumed in pairs. Since Pt efficiently catalyzes O<sub>2</sub> reduction, the hole oxidation of phenol on Pt/TiO<sub>2</sub> is faster than that on TiO<sub>2</sub>, as observed in absence of Ni<sup>2+</sup> (Fig. 1A). The Pt-mediated electron transfer would promote the NiOH<sup>+</sup>-mediated hole transfer, and vice versa. As a result, the efficiency of charge separation is further improved, and the rate increase of phenol degradation on addition of Ni<sup>2+</sup> is much larger with Pt/TiO<sub>2</sub> than that with TiO<sub>2</sub> (Fig. 2B). On the other hand, the amount of O<sub>2</sub> adsorption on TiO<sub>2</sub> in aqueous solution is reduced on addition of Ni<sup>2+</sup>, due to surface occupation of NiOH<sup>+</sup> (Fig. 5). The reduced amount of O<sub>2</sub> on photocatalyst would then result into a decreased rate of O<sub>2</sub> reduction, and hence a decreased rate of phenol degradation. In practice, however, the rate of phenol degradation increased with C<sub>Ni</sub> at low C<sub>Ni</sub> (Fig. 2B). These observations further indicate that the hole oxidation of Ni<sup>2+</sup> is more efficient than the hole oxidation of phenol. At high C<sub>Ni</sub>, the inhibitory effect of Ni<sup>2+</sup> on O<sub>2</sub> adsorption becomes dominant, and the rate of phenol degradation then decreased with C<sub>Ni</sub> (Fig. 2B). Since Pt/TiO<sub>2</sub> is much more active than TiO<sub>2</sub>, the Ni<sup>2+</sup> caused rate decrease of phenol degradation on Pt/TiO<sub>2</sub> is also much larger than that on TiO<sub>2</sub> (Fig. 2B). In other words, the effect of Pt and NiOH<sup>+</sup> mutually cooperate, greatly changing the photocatalytic activity of TiO<sub>2</sub> for O<sub>2</sub> reduction and phenol oxidation, as outlined in Scheme 1.

Furthermore, an alternative mechanism of Ni<sup>0</sup> as cocatalyst has been proposed in the literature [33,36]. For the photocatalytic production of H<sub>2</sub> in presence of excess ethanol, a Ni(OH)<sub>2</sub>-deposited P25 TiO<sub>2</sub> was much more active than P25 TiO<sub>2</sub>. The observed positive effect of Ni(OH)<sub>2</sub> is ascribed to the reduction of Ni(OH)<sub>2</sub> into Ni<sup>0</sup> clusters, followed by the Ni-catalyzed reduction of H<sup>+</sup> to H<sub>2</sub>. In our sample, however, no Ni<sup>0</sup> was observed by XPS (Fig. 3B). To verify this pathway, experiment with an aqueous suspension containing 2 g/L of TiO<sub>2</sub> and 10% of CH<sub>3</sub>OH in volume was carried out under N<sub>2</sub>. After light irradiation for 2 h, the amounts of H<sub>2</sub>, produced from TiO<sub>2</sub> and TiO<sub>2</sub> plus 1 mM Ni<sup>2+</sup>, were 322 and 4506 μmol/g, respectively. Meanwhile, the TiO<sub>2</sub> suspension changed its color from white to blue, but the suspension containing Ni<sup>2+</sup> changed its color from white to grey, with some black particles in the position of a LED lamp (Fig. S9). The former is due to the formation and accumulation of Ti<sup>III</sup> species, because the electron reduction of protons is much slower than the hole oxidation of CH<sub>3</sub>OH. The black particles are probably Ni<sup>0</sup> clusters. In aqueous solution at pH 0, the Ni<sup>2+</sup>/Ni couple has a potential (−0.25 V vs. NHE) more negative than E<sub>CB</sub> of anatase TiO<sub>2</sub> (−0.12 V vs. NHE) [7]. Then in thermodynamics, the reduction of Ni<sup>2+</sup> by the photoelectrons of TiO<sub>2</sub> is impossible. We speculate that the observed black Ni particles result from

the reduction of Ni<sup>2+</sup> by H<sub>2</sub> produced *in situ* under N<sub>2</sub>. In the present work, therefore, a Ni<sup>3+</sup> species formed on the surface of TiO<sub>2</sub> from the hole oxidation of NiOH<sup>+</sup> under air is responsible for the observed enhancement of phenol degradation on addition of Ni<sup>2+</sup>.

#### 4. Conclusions

A simple addition of Ni<sup>2+</sup> ions into the aerated aqueous suspension of TiO<sub>2</sub> can result into great enhancement in the rates of phenol oxidation and mineralization under UV light. A maximum reaction rate is located around initial pH 7, with Ni<sup>2+</sup> ions occupying half the surface of TiO<sub>2</sub> at approximately 50%. Through a theoretical calculation, the observed positive effect of Ni<sup>2+</sup> is ascribed to the strongly adsorbed NiOH<sup>+</sup> ions on TiO<sub>2</sub>, which are more active than the loosely deposited Ni(OH)<sub>2</sub> on TiO<sub>2</sub>. During phenol degradation, Ni<sup>2+</sup> ions act as co-catalyst, without change in concentration. Since Pt efficiently catalyzes O<sub>2</sub> reduction, not only phenol degradation on Pt/TiO<sub>2</sub> is much faster than that on TiO<sub>2</sub>, but also the effect of Ni<sup>2+</sup> ions to Pt/TiO<sub>2</sub> is larger than that to TiO<sub>2</sub>. On the irradiated Pt/TiO<sub>2</sub>, moreover, there is formed a Ni<sup>3+</sup> species, not Ni<sup>0</sup>. With the aid of (photo)electrochemical measurement, it is proposed that NiOOH is formed from the hole oxidation of NiOH<sup>+</sup>, followed by phenol oxidation to regenerate Ni<sup>2+</sup>. Then the NiOH<sup>+</sup>-mediated hole promotes the Pt-mediated electron transfer, and vice versa. As a result, the efficiency of charge separation is improved, and the rate of phenol degradation is increased. Since Ni<sup>2+</sup> ions are often present in water, and they are also cheap in cost, this work has a good prospect of using Ni<sup>2+</sup> ions as the hole mediator of organic oxidation on different semiconductors.

#### Declaration of Competing Interest

None.

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#### Appendix A. Supplementary data

Supplementary material related to this article can be found, in the online version, at doi:<https://doi.org/10.1016/j.apcatb.2019.117903>.

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